

Climate change and decadal shifts in the phenology of larval fishes in the California Current ecosystem

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Climate change has prompted an earlier arrival of spring in numerous ecosystems. It is uncertain whether such changes are occurring in Eastern Boundary Current Upwelling ecosystems, because these regions are subject to natural decadal climate variability, and regional climate models predict seasonal delays in upwelling. To answer this question, the phenology of 43 species of larval fishes was investigated between 1951 and 2008 off southern California. Ordination of the fish community showed earlier phenological progression in more recent years. Thirty-nine percent of seasonal peaks in larval abundance occurred earlier in the year, whereas 18% were delayed. The species whose phenology became earlier were characterized by an offshore, pelagic distribution, whereas species with delayed phenology were more likely to reside in coastal, demersal habitats. Phenological changes were more closely associated with a trend toward earlier warming of surface waters rather than decadal climate cycles, such as the Pacific Decadal Oscillation and North Pacific Gyre Oscillation. Species with long-term advances and delays in phenology reacted similarly to warming at the interannual time scale as demonstrated by responses to the El Niño Southern Oscillation. The trend toward earlier spawning was correlated with changes in sea surface temperature (SST) and mesozooplankton displacement volume, but not coastal upwelling. SST and upwelling were correlated with delays in fish phenology. For species with 20th century advances in phenology, future projections indicate that current trends will continue unabated. The fate of species with delayed phenology is less clear due to differences between Intergovernmental Panel on Climate Change models in projected upwelling trends.

phenology | fish larvae | upwelling ecosystem | California Current | global change biology

Phenology is the study of seasonal biological processes and how they are influenced by climate and weather. Because warmer temperatures are frequently associated with earlier phenological events, changes in phenology are common indicators of the effects of climate change on ecological communities. Meta-analyses have shown that phenological events have advanced at mean rates of 2–5 d/decade relative to a historical baseline (1–5). Among species for which phenological changes were detected, >80% of these shifts occurred in a direction consistent with climate change (1, 2, 5).

Despite the inclusion of hundreds of species in meta-analyses examining climate change effects on phenology, gaps in knowledge persist because most long-term studies of phenology have monitored spring events affecting terrestrial species residing in temperate habitats in the northern hemisphere (2, 4). Marine species are particularly underrepresented in these meta-analyses (6), although see recent work by Poloczanska et al. (5). Underrepresentation of marine species is problematic not only due to their ecological importance, but also because research suggests shifts in phenology may occur more rapidly in marine environments than terrestrial ecosystems (7, 8). Compared with other marine organisms, there is a longer history of studying the phenology of teleost fishes, because the seasonal coincidence between phytoplankton blooms and fish spawning can affect

recruitment to commercial fisheries (9, 10). Nevertheless, little research has investigated the impact of anthropogenic climate change on fish phenology and, with a few exceptions (11, 12), studies have been limited to a small set of commercially fished species. To address this issue, the present study investigated the influence of local and basin-scale climatic and oceanic factors on the phenology of 43 fish species whose larval abundance has been monitored off southern California since 1951.

In the California Current Ecosystem (CCE), wind-driven upwelling is one of the predominant physical processes regulating the seasonal cycle of primary and secondary productivity. The seasonal onset of upwelling, referred to as the “spring transition,” is associated with the commencement of southward, alongshore winds that induce offshore Ekman transport of coastal waters (13). This process coincides with a decrease in coastal sea surface temperature (SST) by 1.5–4 °C, the development of a nearshore, southward oceanic jet, and an increase in chlorophyll *a* in the coastal zone (14). In the central and northern CCE (>35°N), upwelling intensifies through summer until the wind direction reverses in the fall. In the southern CCE (<35°N), upwelling is observed in all seasons, but its intensity diminishes in fall and winter (15, 16). This pattern leads to a spring maximum in chlorophyll concentration occurring between March and May in the southern CCE (17, 18). The peak in phytoplankton is followed by a summer maximum in mesozooplankton displacement volume between May and July (19). Fishes spawn in the southern CCE year round, although most species exhibit distinct seasonal patterns of larval abundance (20).

There is increasing evidence that seasonal cycles of temperature, sea surface height (SSH), and chlorophyll concentration may

Significance

In terrestrial ecosystems, earlier phenology (i.e., seasonal timing) is a hallmark organismal response to global warming. Less is known about marine phenological responses to climate change, especially in Eastern Boundary Current Upwelling (EBCU) ecosystems that generate >20% of fish catch. The phenology of 43 EBCU fish species was examined over 58 years; 39% of phenological events occurred earlier in recent decades, with faster changes than many terrestrial ecosystems. Zooplankton did not shift their phenology synchronously with most fishes. Fishes that aren't changing their phenology synchronously with zooplankton may be subject to mismatches with prey, potentially leading to reduced recruitment to fisheries. Adjusting the timing of seasonal management tactics (e.g., fishery closures, hatchery releases) may help ensure that management remains effective.

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not be stationary in the CCE (21, 22). Empirical observations and regional climate models suggest that climate change is leading to intensification of upwelling during spring and/or summer months but not other seasons (15, 23, 24). A model simulation that doubled atmospheric CO₂ indicated that upwelling off northern California is likely to increase during July–October, but decrease in April–May, resulting in a 1-mo delay in the spring transition (15). When climate feedbacks due to changes in land cover were accounted for in this model, similar results were obtained for the northern CCE, but model predictions suggested that the southern CCE would experience an increase in early-season upwelling and decreased peak and late-season upwelling (24). This change could cause an advance in the seasonal cycle of southern CCE upwelling, as well as potentially dampen its seasonal amplitude. Empirical observations of ocean temperature and the Bakun upwelling index validate these model results, indicating delays in the onset and peak of upwelling, particularly in the northern CCE (21, 25). In accordance with model predictions of earlier upwelling in the southern CCE, phytoplankton blooms were observed 1–2 mo earlier in the late 1990s in this region compared with previous years (18).

In addition to anthropogenic climate change, oceanography in the CCE is affected by climate oscillations with interannual-to-decadal periods, including El Niño–Southern Oscillation (ENSO), the Pacific Decadal Oscillation (PDO), and the North Pacific Gyre Oscillation (NPGO). El Niño is frequently accompanied by delays in seasonal upwelling (25), although lower frequency modes of climate variability, such as the PDO, do not have as pronounced an effect on upwelling seasonality (19, 25). Due to delayed and reduced upwelling, El Niño is associated with later spring phytoplankton blooms in certain regions (26). Although little research has examined PDO effects on bloom timing in the CCE, this mode of climate variability influences phytoplankton phenology elsewhere in the North Pacific (27). Among zooplankton in the CCE, the 1977 change from a negative to a positive PDO coincided with a 2-mo shift toward earlier maximum displacement volume of zooplankton (19). At the next trophic level, El Niño influences the spawning phenology of northern anchovy (*Engraulis mordax*) (28). Also, early spawning migrations of chinook salmon (*Oncorhynchus tshawytscha*) are weakly correlated with warm PDO anomalies (29). The NPGO is a more recently defined climate oscillation based on the second mode of SSH variability in the Northeast Pacific (30). In the southern CCE, the NPGO is more closely correlated to variations in upwelling, salinity, nutrients, and chlorophyll than the PDO. Whether this mode of climate variability has an impact on the phenology of marine organisms is a subject yet to be investigated.

The match-mismatch hypothesis provides a mechanism explaining how climate-induced changes in fish and plankton phenology could potentially alter the abundance of fish stocks (9). This hypothesis proposes that fishes spawn during peak seasonal plankton production, which increases the likelihood that their larvae will encounter sufficient prey. However, due to interannual variability in the timing of plankton production and fish spawning, these events do not always coincide. During such mismatches, first-feeding larvae may experience increased vulnerability to starvation or slower growth, which can heighten susceptibility to predation (31). Poor larval survival can result in reduced recruitment and decreased fishery landings in subsequent years. Although many other processes during the early life history of fishes influence recruitment (32), variations in plankton phenology can result in order-of-magnitude changes in the recruitment and survival of commercially important fishes (33, 34). Coastal upwelling in the CCE complicates mismatch dynamics, because upwelling simultaneously provides nutrients for planktonic production while advecting fish larvae away from coastal habitats. Due to these opposing influences on larval survival and growth, many fishes in

Eastern Boundary Current Upwelling (EBCU) systems spawn at low-to-moderate rates of upwelling (9, 35).

Climate change could increase phenological mismatches through two mechanisms. First, certain seasonal cues, such as day length, will not be affected by global warming, whereas other seasonal processes will exhibit differing rates of change (e.g., surface vs. bottom temperatures) (36, 37). Because predators and prey may use different environmental factors as signals to initiate seasonal behaviors, this discrepancy can lead to a higher frequency of mismatches if these indicators become decoupled (10, 36). Second, individual species inevitably have distinct climate sensitivities that result in differing rates of phenological response to climate change. These differences can lead to situations where even small changes in climate can upset seasonal, interspecific interactions (10).

This study examined decadal changes in the phenology of 43 species of fish larvae off California. The objectives were to (i) determine whether phenological changes were correlated with variations in regional climate indices and the seasonality of SST, coastal upwelling, and zooplankton displacement volume; (ii) assess whether life history characteristics of fishes are linked to changes in phenology, and (iii) forecast 21st century changes in fish phenology based on predicted changes in seasonal SST and coastal upwelling.

Several hypotheses can be proposed regarding how fish phenology has changed since 1951 when ichthyoplankton surveys began in the southern CCE:

- H₀: The phenology of larval fishes will remain constant regardless of variations in seasonal oceanographic conditions.
- H_{A1}: Fish larvae will uniformly occur earlier during periods with warmer temperature, reflecting earlier spawning. This change could arise due to accelerated oocyte development under warmer temperatures (38–40) or the tendency of spawning to track phytoplankton blooms, which have occurred earlier in the southern CCE in recent years (18).
- H_{A2}: Spring-spawning fishes will exhibit earlier larval phenology during periods with warmer temperatures, whereas fall-spawning species will exhibit delayed phenology, reflecting a later onset of cooler, fall conditions.
- H_{A3}: Delays in upwelling will lead to later spawning and occurrence of larvae.
- H_{A4}: The phenology of larval fishes will display interannual-to-decadal variability synchronous with climate oscillations, such as ENSO, PDO, and/or NPGO.

To examine these hypotheses, data on the abundance of larval fishes from California Cooperative Oceanic Fisheries Investigations (CalCOFI) were used. CalCOFI has surveyed larvae between 1951 and 2008 on a monthly-to-quarterly basis (with some gaps between 1967 and 1983) (20, 41). The region most consistently surveyed by CalCOFI includes the Southern California Bight (SCB), the area offshore of the SCB, and Point Conception (Fig. S1).

Monthly abundance of 43 fish species was calculated by decadal averaging data from quarterly surveys conducted in different months in successive years. This step was undertaken to achieve the minimum of a monthly sampling resolution needed for examining phenological trends. Eight species exhibited two peaks in larval abundance each year (Table S1). The peaks were classified as separate phenological events, or phenophases, because each peak for a given species may display a different trend in phenology. Changes in phenology were detected based on the central tendency (CT) of seasonal larvae occurrence each decade (7, 12, 42). Because calculating CT relative to decadal means led to a small sample size for each phenophase ($n = 6$) and reduced statistical power, this study did not focus on species-level changes in phenology. Instead, each phenophase was treated as a replicate for examining assemblage-wide patterns. Variations in CT

were treated as a proxy for spawning time, because CalCOFI mainly collects young, preflexion larvae (43, 44), and the egg stage of many species lasts 2–4 d at temperatures in the southern CCE (45). Depending on the species and temperature, flexion occurs between 3 and 25 d after hatching (45–47).

Results

Detection of Phenology Trends. The first principal component (PC) of the phenology of the larval fish assemblage indicated that larvae are appearing progressively earlier in the year (Fig. 1). This PC accounted for 32.6% of the variance in CT anomalies calculated relative to each species' multidecadal mean. Similarly, linear regression of CT anomalies vs. year showed that on average fishes have advanced their phenology by 7.2 d since the 1950s ($Y = 9.3305 - 0.0047X$, $F = 6.5$, $df = 288$, $P < 0.05$). CT anomalies suggested that this trend is mainly due to changes over the last two decades (Fig. S2C).

Because divergent trends among species would not be evident when examining the assemblage mean, most ensuing analyses were performed on three groups of phenophases that shared similar decadal patterns. This categorization was based on the correlation between the CT of each phenophase and time, where phenophases with a correlation coefficient $r \geq 0.5$ were classified as having later phenology, $r \leq -0.5$ indicated earlier phenology, and intermediate values corresponded to phenophases with no long-term, linear changes. Twenty phenophases fell into the earlier phenology group; 9 were classified as exhibiting later phenology, and 22 did not show a trend (Table S2). Regressions of CT anomalies vs. decade for each group demonstrated that changes were statistically significant for earlier and later groups (earlier species: $Y = 53.4012 - 0.0269X_1 - 0.4070X_2$, where $X_1 = \text{decade}$ and $X_2 = \text{first-order autoregressive term}$, $F = 33.6$, $df = 87$, $P < 0.001$; later species: $Y = -40.7386 + 0.0206X_1 - 0.3214X_2$, $F = 10.4$, $df = 40$, $P < 0.001$). Among species with earlier phenology, the rate of phenological change varied between -2.8 to -12.4 d/decade, with a mean of -6.4 d/decade (Table S2). Commercially fished species with earlier phenology included jack mackerel (*Trachurus symmetricus*), Pacific hake (*Merluccius productus*), Pacific mackerel (*Scomber japonicus*), and three rockfishes (*Sebastes aurora*, *S. diploproa*, and *S. jordani*).

Species with later phenology shifted their spawning times at a slightly slower rate (mean: 5.1 d/decade; range: 3.0–7.2 d/decade; Table S2); than the earlier phenology group. The time series of CT anomalies for later species showed an abrupt shift between the 1970s and 1980s (Fig. S2I). Species with later phenology that are targeted by commercial fisheries included chilipepper (*Sebastes goodei*) and the spring phenophase of California halibut (*Paralichthys californicus*).

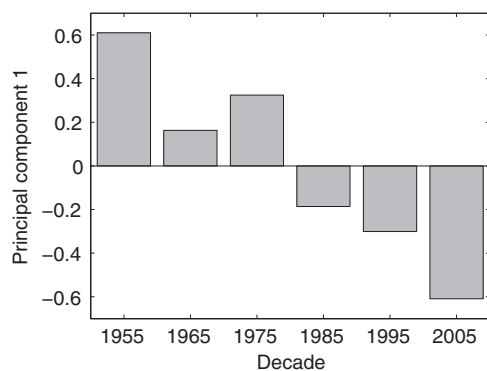


Fig. 1. First principal component of the central tendency of seasonal occurrence of larval fishes. Positive values of eigenvectors on the y axis indicate later occurrence of larvae, whereas negative values indicate earlier occurrence.

A Monte Carlo simulation was used to determine the detection limits of phenological change given the CalCOFI survey design (SI Text and Tables S3 and S4). Based on this simulation, the 90% CI of changes in CT was estimated to be ± 2.8 – 3.1 d/decade. These CIs corresponded closely to the minimum rate of change observed among species in the earlier and later phenology groups.

For species classified as having no long-term trend in phenology, their mean rate of phenological change was -0.4 ± 1.0 SE d/decade (Table S2). However, some species in this group exhibited large, interdecadal fluctuations in spawning times. For example, the CT of *Sardinops sagax* had an SD of 67.2 d. Other commercially important species in this group were northern anchovy (*E. mordax*), California halibut (*P. californicus*, summer phenophase), Pacific sanddab (*Citharichthys sordidus*, both phenophases), English sole (*Parophrys vetulus*), and bocaccio (*Sebastes paucispinis*).

It is important to examine whether the detected phenological trends could be due to decadal variations in the intensity of CalCOFI sampling. During periods of greater sampling effort, there is a higher likelihood of detecting precocious individuals who spawn early even if no change in the mean distribution of spawning occurred (48). CalCOFI survey effort was greatest in the 1950s when 3,952 ichthyoplankton samples were collected in the southern CCE. Effort dropped during the 1960s and 1970s, such that 1,614 samples were collected during the latter decade. Following the resumption of annual sampling in 1984, between 2,228 and 2,651 samples per decade were collected. As a result, trends in phenology cannot be attributed to a long-term increase or decrease in sampling effort. Similarly, the increasingly standardized timing of cruises, which led to gaps in monthly sampling during the 2000s, did not have a major effect on the first PC of fish phenology nor the three phenology groups (Fig. S2 and SI Text).

Relationships Between Phenology, Climate, Oceanic Conditions, and Ecological Traits. Fish phenology changed markedly during El Niño events, but the direction of changes depended on phenology group. For species that exhibited a decadal shift toward earlier phenology, El Niño (La Niña) was also associated with earlier (later) occurrence of larvae (Fig. 2A; Kruskal–Wallis, $\chi^2 = 7.5$, $df = 2,57$, $P < 0.05$). This result suggested that fishes in this group reacted similarly to warming at the interannual scale associated with ENSO and the decadal scale. In contrast, species in the two other phenology groups experienced delays (advances) in spawning during El Niño (La Niña) (Fig. 2B and C; species without a linear trend: $F_{2,63} = 6.3$, $P < 0.01$; later species: $F_{2,24} = 5.3$, $P < 0.05$). This response may reflect a late onset of upwelling and primary production during El Niño in the CCE (26).

Attributing changes in phenology to the PDO is complicated by the fact that CalCOFI started during a negative cold PDO period that persisted through 1976 and was followed by a positive warm PDO phase continuing through 1998. Before 1998, the signal of anthropogenic climate change and warming related to the PDO were confounded. The return of negative PDO conditions in 1999 may potentially allow for clearer distinction between these influences on phenology (49). If the PDO was associated with changes in phenology, then the timing of larval occurrence during recent, negative PDO years should be similar to that observed during the earlier, negative PDO. To test for this effect, ANOVAs were performed comparing fish phenology during the first negative PDO phase (1951–1976), the second negative PDO phase (1999–2002 and 2007–2008), and the warm PDO phase (1977–1998; Fig. 2D–F). These ANOVAs indicated that there were changes in phenology across these periods for both species with long-term delays and advances in phenology (earlier species: $F_{2,57} = 21.5$, $P < 0.001$; later species: $F_{2,24} = 11.4$, $P < 0.001$). No significant effects were observed for species without trends in phenology ($F_{2,63} = 0.1$, $P = 0.88$). For advancing

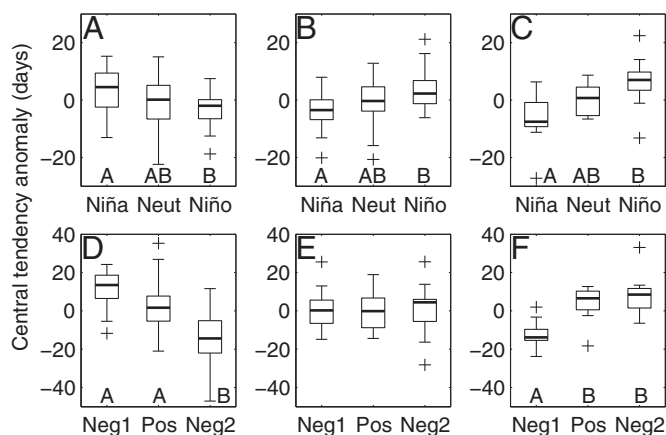


Fig. 2. Effects of El Niño-Southern Oscillation (A–C) and Pacific Decadal Oscillation (PDO; D–F) on species displaying earlier phenology (Left; $n = 60$), no long-term, linear change in phenology (Center; $n = 66$), and later phenology (Right; $n = 27$). In each box plot, the darkened line indicates the median; boxes show the interquartile range; whiskers indicate the expected extent of 99% of the data for a Gaussian distribution, and crosses show outliers. Two outliers fall outside the range displayed in A. Capital letters below boxes indicate means that differ significantly at $P < 0.05$ based on Tukey–Kramer multiple comparison tests. (A–C) El Niño, neutral (Neut), and La Niña conditions based on the Oceanic Niño Index. (D–F) Neg1, cold-phase PDO between 1951 and 1976; Pos, warm-phase PDO from 1977 to 1998; Neg2, cold-phase PDO during 1999–2002 and 2007–2008.

species, Tukey–Kramer multiple comparison tests revealed that these differences were due to changes in phenology between the recent, negative PDO and previous years (Fig. 2D). Phenology during the first, negative PDO differed from all subsequent periods among species with later phenology (Fig. 2F). In neither case were the first and second negative PDO periods similar to each other.

A similar analysis was conducted to evaluate whether fish phenology was influenced by negative, neutral, or positive NPGO conditions. ANOVAs testing this association were not significant for any phenology group (Fig. S3). This result again suggested that assemblage-wide changes in phenology may be more indicative of long-term trends than decadal climate variability.

To further investigate which factors were responsible for changes in fish phenology, the effects of three locally measured oceanic variables (e.g., CalCOFI SST, CalCOFI zooplankton displacement volume, and the Bakun upwelling index) were examined. CT anomalies of these variables were calculated and regressed against time to assess decadal changes in seasonal patterns. Changes in SST CT indicated that ocean temperatures are now warming 25.9 d earlier than in the 1950s (Fig. 3A; $Y = 33.7200 - 0.0170X$, $F = 11.0$, $df = 4$, $P < 0.05$). Upwelling did not exhibit a significant trend in decadal phenology (Fig. S4A; $F = 1.2$, $df = 4$, $P = 0.34$), although peak upwelling during the 1980s and 2000s occurred slightly later than at the start of the time series. The CT of zooplankton volume was characterized by delayed phenology in the 1970s and early phenology in the 1980s (Fig. S4B), possibly reflecting changes related to the 1977 PDO shift (19). CT anomalies of zooplankton volume were close to zero in other decades, with no long-term trend ($F = 1.0$, $df = 4$, $P = 0.30$).

Species in the earlier phenology group displayed significant, positive correlations with the CT of SST and zooplankton volume but were not influenced by upwelling (Fig. 3 and Table S5). Species whose phenology was delayed since the 1950s spawned later when the onset of upwelling also occurred later (Fig. 3D). In addition, this phenology group was inversely correlated with SST CT, such that fishes spawned earlier in years when seasonal temperatures remained cool for a longer time (Fig. 3B). CT

anomalies for species with no long-term trends in phenology were not correlated with any oceanic variables (Table S5).

Next, this study investigated whether species that displayed phenological changes shared certain ecological traits. The characteristics examined were the six most common taxonomic orders, season of maximum larval abundance, the amplitude of a species' seasonal cycle, trophic level, habitat of adult fishes, cross-shore distribution of larvae, biogeographic affinity (e.g., warm water, cool water, or widespread distribution), frequency of larval occurrence during surveys, and fishing status (Tables S1 and S2). The predominant modes of variability among these traits were characterized with PC analysis (PCA). The first two PCs constituted 45.5% of the variance in this dataset. An ordination of fishes with respect to these two PCs revealed three clusters of species with distinct traits (Fig. 4A). The first cluster was associated with positive loadings on PC1 and primarily negative loadings on PC2. Fishes in this group were mainly mesoepelagic with an oceanic distribution (Fig. 4C). Myctophiformes and Stomiiformes were common in this cluster (Fig. 4B). The second cluster contained species with negative loadings on both PC1 and PC2. The distinguishing features of this cluster were that fishes used coastal, demersal habitats with cool water masses. Scorpaeniformes and Pleuroniformes were the two characteristic orders of fishes in this cluster. The third cluster indicated by positive loadings on PC2 was comprised of epipelagic fishes with a widespread, coastal-oceanic distribution, a high frequency of occurrence, and a large, seasonal amplitude (Fig. 4C and D). Many fishes in this cluster belonged to the order Perciformes.

Both PC1 and PC2 were negatively correlated with changes in CT ($r = -0.25$ and -0.31 , respectively). This relationship was

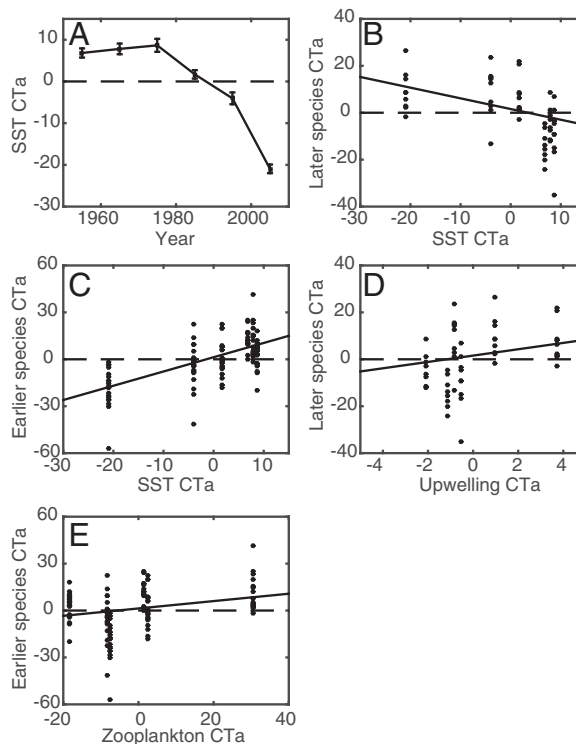


Fig. 3. Responses of the phenology of larval fishes to changes in the central tendency (CT) of SST, zooplankton displacement volume, and the Bakun upwelling index at 33°N and 119°W. Units of all CT anomalies (CTa) are days. (A) Decadal changes in SST CT. SEs from bootstrap analysis are shown. (B–E) Multiple regression equations shown in Table S5. Variables included in a given regression, but not displayed on these bivariate plots, were held constant at a CT anomaly of zero.

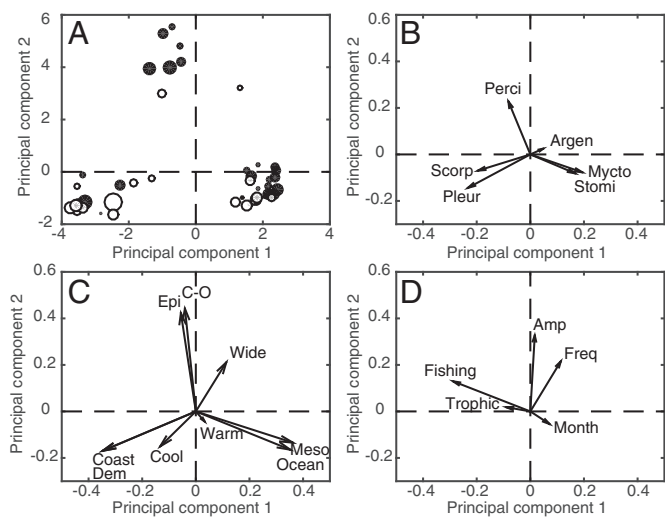


Fig. 4. Principal components analysis performed on ecological traits ($n = 51$). (A) Species ordination. Marker size is proportional to changes in CT. Black (white) dots indicate species whose CT has a negative (positive) slope when regressed against year. (B) Eigenvectors of taxonomic orders. Stomi, Stomiiformes; Perci, Perciformes; Scorp, Scorpaeniformes; Mycto, Myctophiformes; Argen, Argentiniformes; Pleur, Pleuronectiformes. (C) Eigenvectors of adult fish habitat (epipelagic, mesopelagic, demersal), cross-shore distribution [coastal, coastal-oceanic (C-O), oceanic], and biogeographic affinity (cool water, warm water, widespread). (D) Eigenvectors of adult trophic level, fishing status, month of maximum larval abundance, frequency of occurrence, and amplitude of the seasonal cycle.

significant at $P < 0.05$ for PC2, but marginally insignificant ($P = 0.07$) for PC1. Consequently, coastal-oceanic, epipelagic species in the third cluster mainly exhibited phenological advances, with a mean rate of change of -4.3 d/decade ($Y = 28.2158 - 0.0142X_1$, where $Y =$ CT anomalies and $X_1 =$ year, $F = 3.9$, $df = 46$, $P = 0.05$, Fig. 4A). Advancing phenology was also observed among 70% of oceanic, mesopelagic species in the first cluster. The rate of phenological change for this cluster was -3.0 d/decade ($Y = 19.7065 - 0.0099X_1 - 0.2294X_2$, where $X_2 =$ first-order, autoregressive term, $F = 11.4$, $df = 117$, $P < 0.001$). Delayed phenology was detected most frequently among coastal, demersal fishes in the second cluster, although this group's rate of phenological change was not statistically significant ($F = 0.4$, $df = 93$, $P = 0.54$). Based on a comparison between the current study and changes in the geographic distribution of the same fish species (50, 51), one additional ecological characteristic of the earlier phenology group was that they exhibited a higher probability of undergoing range shifts in response to changes in climatic and oceanic conditions (one-tailed binomial test: $n = 16$, $P < 0.05$).

Projections of 21st Century Changes in Phenology. The Intergovernmental Panel on Climate Change (IPCC) Special Report on Emissions Scenarios (SRES) A1B scenario was used to forecast future changes in fish phenology. SRES A1B is a medium-range emissions scenario, where atmospheric CO_2 stabilizes at 720 ppm (52). Projections were based on the empirical relationship between local, oceanic variables and species with earlier or delayed phenology (Fig. 3 and Table S5). Changes in mean phenology between 2000–2009 and 2090–2099 were forecasted with 12 IPCC models run over an ensemble of 30 simulations. Models unanimously projected that SST would continue to warm earlier in the year through the 21st century. Predicted changes in SST CT between the 2000s and 2090s ranged between -1.5 and -18.0 d (mean = -10.1 d). In contrast, model outputs indicated high uncertainty in future changes in upwelling seasonality in the southern CCE, with 15 model runs predicting earlier upwelling

and 11 predicting later upwelling. Mean change in upwelling CT between 2000–2009 and 2090–2099 was -4.4 d, with a range of -18.9 to 11.8 d. Projections for species with earlier phenology indicated that fishes would advance their phenology by -5.7 d on average relative to current conditions (Fig. 5; range, -0.1 to -10.4 d). Due to inconsistencies between models when forecasting changes in upwelling CT, projections suggested that species with later phenology may spawn anywhere between 21.6 d earlier to 22.1 d later compared with current conditions (Fig. 5B; mean = -1.7 d).

Discussion

Evaluation of Hypotheses. This study indicated that the reproductive phenology of marine fishes showed varying responses to recent climate change that could be explained in part by the habitat used by each species. Despite the presence of strong natural climate variability in the CCE, long-term trends were detected in the phenology of 26 out of 43 fish species. Because trends in upwelling and zooplankton volume were not evident over this period, fishes with shifting phenology potentially could be subject to seasonal mismatches with lower trophic levels.

At the outset of this study, several hypotheses were proposed regarding phenological changes. The null hypothesis (H_0) stated that fish phenology would remain nearly constant through time. This outcome would be expected if a biological or environmental variable that exhibits little decadal variability were the primary influence on the timing of fish reproduction. For example, photoperiod is often cited as a process that affects the initiation of fish reproductive development (53) and displays little inter-annual variability. Although 22 fish phenophases (43%) showed no long-term trends, as is consistent with H_0 , 29 phenophases (57%) displayed a decadal trend in phenology, suggesting a key

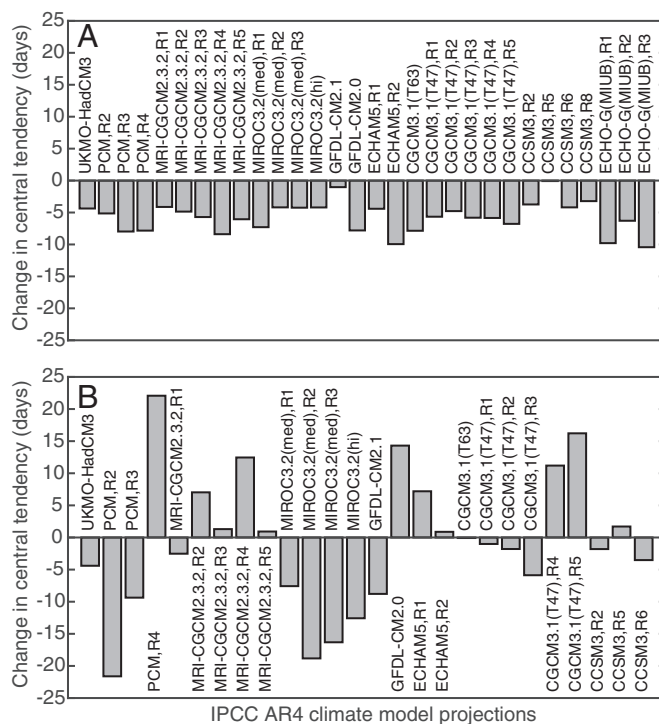


Fig. 5. Projected changes in the phenology of larval fishes between 2000–2009 and 2090–2099 based on climate models from the IPCC Fourth Assessment Report. All projections are based on IPCC scenario A1B. Abbreviated IPCC names for each model are shown. (A) Species whose phenology became earlier during the 20th century. (B) Species whose phenology became later during the 20th century.

role of climate. Furthermore, even species in the phenology group with no long-term trend exhibited interannual variations in larval occurrence associated with ENSO (Fig. 2B), indicating that their phenology was not as unwavering as suggested by H_0 . Also, some species in this phenology group, such as *S. sagax*, displayed large interdecadal fluctuations in their phenology in absence of a long-term trend.

H_{A1} proposed that earlier occurrence of larvae would be observed during periods with warm temperatures. Some support for H_{A1} was provided by the positive correlation between SST CT and species with earlier phenology (Fig. 3C). Similar relationships between temperature and fish phenology were identified among ichthyoplankton in the North Sea and English Channel (11, 12). However, these studies considered a much smaller area (one to three sites), and their time series were not long enough to differentiate between the effects of climate change and climate variability.

Although the relationship identified between fish phenology and SST CT is correlative rather than mechanistic, several lines of evidence buttress this conclusion. First, temperature has a large physiological effect on fish reproduction due to its influence on the production of hormones, vitellogenin, oocyte development, and egg survival (53). Because temperature accelerates gonadal maturation in poikilothermic fishes, changes in seasonal temperature measured in degree days has been used to forecast spawning time in diverse taxa (38–40). Second, species in each phenology group reacted similarly to decadal warming and changes in SST associated with ENSO indicating a consistent response of phenology to temperature across different time scales.

The effect of SST CT on larval phenology could reflect in part temperature-induced changes in the duration of the egg and larval stages among poikilotherms. Although a shortened duration of these life history stages under warmer temperatures could be partially responsible for phenological advances, it is unlikely that changes in stage duration alone could result in the observed rates of 3–12 d/decade of phenological change. The egg stage of many fishes only lasts 2–4 d at temperatures in the southern CCE (45). Also, larger, late-stage larvae that would be more affected by changes in stage duration were rare in CalCOFI samples (54). In addition to direct physiological effects, the influence of temperature on phenology may be mediated through other biological processes, such as seasonal migrations of fishes through regions with differing temperatures (12, 42) and the influence of temperature on seasonal prey availability (55).

Two hypotheses were proposed to explain delays in fish phenology. H_{A2} suggested that fishes spawning in fall may exhibit delayed phenology if climate change leads to extended warm summer temperatures, whereas H_{A3} asserted that delayed phenology would result from later onset of upwelling. This study provided scant support for H_{A2} , because the season of maximal larval abundance did not have a sizable effect on trends in fish phenology (Fig. 4D). Only two of the nine phenophases with delayed phenology were fall-spawning fishes. Although a long-term trend in upwelling seasonality was not observed (Fig. S4A), species with delayed phenology showed a positive correlation with upwelling CT (Fig. 3D), supporting H_{A3} . Upwelling could influence fish phenology through either offshore transport of larvae or augmentation of primary production. Among coastal, demersal fishes that were more likely to show delayed phenology (Fig. 4), it is imperative to minimize advection of larvae into unsuitable, offshore habitats. Because upwelling results in cross-shore transport, many coastal fishes spawn during winter when upwelling is minimal to increase retention (35). The phenology of these species may be positively correlated with the timing of upwelling to maintain their reproductive activity in synchrony with the winter lull in upwelling. For species whose larvae are most abundant in spring and summer, upwelling likely affects phenology through its impact of primary and secondary production, which increases prey availability.

H_{A4} proposed that, in lieu of long-term trends, interannual-to-decadal variation in phenology associated with basin-scale climate cycles would be observed. Although ENSO was related to significant fluctuations in fish phenology, the PDO and NPGO did not have clear-cut effects (Fig. 2 and Fig. S3). This finding may initially seem surprising because the PDO and NPGO affect the abundance and geographic distribution of several species considered here (20, 44, 50, 51, 56). However, time series of CCE ocean temperature (a key factor controlling fish phenology) indicate that, although there have been PDO-like oscillations in the magnitude of this variable, there was also a concurrent, continuous shift in the phase of its seasonal cycle relative to a 1950s baseline (Fig. 3A) (21). This long-term phase shift may explain why the seasonality of larval fishes is less sensitive to the PDO. This result may be part of a more generalized pattern, because the PDO has a weak effect on the phenology of spring chinook salmon (*O. tshawytscha*) (29), a species whose abundance is strongly influenced by this climate cycle (57).

Rates of Current and Future Phenological Change. In the southern CCE, fishes that spawn earlier during warm conditions advanced their phenology at a mean rate of 6.4 d/decade, whereas other species experienced mean delays of 5.1 d/decade. Assuming that temperature is a primary factor affecting fish phenology, the mean temperature sensitivity would be 24.6 d/°C for the earlier phenophases (range: 10.8–47.7 d/°C) and 19.6 d/°C for later phenophases (range: 11.5–27.7 d/°C), because this region warmed by 1.3 °C between 1949 and 2000 (58). When examining all phenology groups jointly using absolute values of phenological change, a mean of 4.7 ± 3.3 SD d/decade or 17.9 ± 12.8 SD d/°C was obtained. These rates were comparable to results from other studies of the phenology of marine and freshwater fishes and meroplankton (5, 7, 12, 40, 42, 59), but were faster than rates from many terrestrial ecosystems (1, 3, 4, 60). Although a sensitivity analysis suggested that this study may have limited power to detect shifts smaller than ~ 3 d/decade (Table S3 and *SI Text*), it is unlikely that the phenological rates calculated here were biased fast because center of gravity metrics, such as CT, are usually conservative and change more slowly than the start date of phenological events (48, 61). Instead, the rapid changes observed in marine systems are more likely due to the smaller amplitude of seasonal temperature changes in marine habitats. This weak seasonal gradient implies that species need to undergo large phenological shifts to ensure that temperature-sensitive activities continue to occur at a constant temperature (8). Furthermore, the high fecundity and dispersal capacity of many fish species may allow them to adapt more rapidly than terrestrial organisms to changing conditions (6).

Another way in which these results differed from previous studies was that a lower percentage of phenophases responded to warming by occurring earlier. In the southern CCE, 39% of phenophases advanced their phenology. In contrast, >60% of species have responded to warming temperatures by advancing their phenology in many terrestrial (1, 3, 60) and aquatic systems (5, 11, 62). Studies of the changing geographic distribution of fishes in the southern CCE also showed that slightly less than half of the species were responsive to climate forcing (50, 51), suggesting that the large percentage of nonresponsive species may be characteristic of the southern CCE and possibly other EBCU ecosystems. This reduced phenological sensitivity to climate may be due to weak seasonality in the southern CCE, its high proportion of mesopelagic fishes that may be less exposed to variations in upper ocean climate (44, 56), or opposing trends in SST and upwelling seasonality.

In addition to long-term phenological advances, 18% of fish phenophases exhibited delayed phenology in the CCE. This percentage is higher than the equivalent percentage in many terrestrial habitats (1, 60) but is comparable to other subtropical,

oceanic regions where delayed zooplankton phenology has been observed (63).

Future projections indicated that fishes in the earlier phenology group would advance their phenology at a mean rate of -0.6 d/decade (range: 0.01 – 1.2 d/decade), which is slower than their current rate of change. For the later phenology group, predictions were highly variable between model runs (Fig. 5B). Both the varying projections for delayed species and the slower rate of predicted change for advancing species may be related to the relatively low spatial resolution ($\geq 1^\circ$) of most IPCC general circulation models (GCMs). This resolution limits the ability of models to represent fine-scale oceanographic dynamics associated with coastal upwelling (64, 65). As a result, most GCMs overestimate temperatures in EBCU systems (65). Overestimation of present day temperatures could potentially lead to an underestimation of future changes in temperature and phenology in these regions. Also, the coarse resolution of GCMs is likely related to the high variability between modeled upwelling rates observed here and in Wang et al. (64). Future predictions of fish phenology may be improved by either using regional climate models or the next generation of global GCMs, some of which will have resolutions as fine as $0.1^\circ \times 0.1^\circ$. Using the ensemble of IPCC-class models from the recent Coupled Model Intercomparison Project–Phase 5 (CMIP5) is unlikely to substantially affect these results because CMIP5 models have a similar spatial resolution to the models examined here and show a similar ensemble range of projected temperature changes over Western North America (52, 66). Even as higher-resolution GCMs become available, the accuracy of projections may be limited if there are thresholds beyond which fishes can no longer adapt to changing conditions by altering their phenology (67). Projections of changes in future fish phenology could also be improved by considering species range shifts, acclimation to changing oceanic conditions, and adaptation through natural selection.

Implications for Marine Ecology and Fisheries Management. Only species in the earlier phenology group were influenced by variations in the seasonality of zooplankton volume. Even in this phenology group, the correlation between the CT of zooplankton and fishes only became evident after fluctuations in SST were taken into account. The diverging phenological trajectories of zooplankton and many fish species suggests that climate change may lead to an increased frequency of mismatches between trophic levels. During the early life history of fishes, mismatches can result in decreased foraging efficiency, slower growth, starvation, and lower survival (10, 34, 68). These impacts may eventually lead to poor recruitment, declines in abundance, and even local extinctions (33, 69), unless rapid evolution reverses phenological trends or density-dependent compensation occurs at a later life stage (70, 71). At times, mismatches between fishes and zooplankton can be transmitted up the food web to higher trophic levels (59).

Nevertheless, zooplankton displacement volume is a bulk measure of a myriad of species, including both predators and prey of larval fishes. Gelatinous organisms (e.g., cnidarians, ctenophores, pelagic tunicates) have a disproportionately large effect on CalCOFI measurements of zooplankton volume (72). Due to the low carbon content of gelatinous zooplankton, displacement volume may be a less than ideal indicator of prey availability for larval fishes. Monitoring of the phenology of individual zooplankton species or functional groups is needed to confirm whether mismatches between fishes and their prey are truly becoming more frequent.

Species that did not exhibit trends in phenology may also be vulnerable to global warming due to their reduced capacity to adapt to changing conditions. A lack of adaptive phenological change in some species has been associated with declining abundance (73). One additional indication of the vulnerability of species in the phenology group with no long-term trend is that

their duration of peak larval abundance has contracted over time (74). A shorter spawning season increases the likelihood that in some years there will be little-to-no temporal overlap between larvae and environmental conditions conducive to their growth and survival. As a result, contracted spawning periods can be accompanied by greater recruitment variability (75).

These factors suggest that phenological plasticity may be a useful metric for predicting which species may benefit from or be impaired by climate change (73). Further evidence of the importance of phenological plasticity comes from the fact that fishes with earlier spawning were more likely to shift their geographic range in response to climatic variations (50, 51), suggesting a link between these two modes of adapting to climate change. Commonalities between species in each phenology group may prove useful for discerning how species yet to be studied will react to climate change. Characteristics that most strongly affected phenological trends were the habitats used by larval and adult fishes (Fig. 4). Taxonomy, relative abundance, and the amplitude of a species' seasonal cycle influenced phenological trends to a lesser extent. Habitat not only influenced phenological trends, but is also a major factor affecting synchronous variations in fish abundance in the CCE (56). Epipelagic species using coastal-oceanic habitats were the group most likely to exhibit earlier spawning. Their phenological plasticity may reflect the fact that many of these species are batch spawners that can reproduce as frequently as once a week (76). These species also live in the upper water column where they are exposed to larger changes in climate. Last, epipelagic species typically have short life spans allowing for possible genetic adaptation to changing conditions. Other studies of climate change effects on fishes have indicated that pelagic species possess heightened climate sensitivity (77).

Current and future changes in fish phenology may necessitate that fishery managers implement precautionary and adaptive management approaches. Management activities that may be affected by changing phenology include (i) seasonal fishing closures designed to protect sensitive life stages or maintain sustainable escapement, (ii) stock assessment surveys timed to coincide with seasonal migrations or the occurrence of a key life history stage (e.g., daily egg production method surveys), and (iii) optimization of the release time of hatchery reared fish for stock enhancement (68). Monitoring long-term changes in phenology and periodically adjusting the timing of seasonal closures, surveys, and hatchery releases will be important for ensuring that management tactics remain effective. In cases where fish phenology is linked to cumulative, seasonal changes in temperature or other variables, it may be possible to develop short-term, operational forecasts to predict when a fishery should open or when a stock assessment survey should occur (78). In cases where climate change leads to increased phenological mismatches and resulting recruitment failure, the carrying capacity of an ecosystem may be reduced, necessitating changes to harvest guidelines. Longer rebuilding times of depleted stocks may also be needed if mismatches lead to prolonged periods of poor recruitment (79).

In conclusion, long-term changes in the spawning phenology of fishes were apparent in the CCE even though this region is strongly influenced by decadal-scale, natural climate variability. Among species whose phenology has become earlier since the 1950s, the mean rate of change was 6.4 d/decade. This rate was much faster than rates from many terrestrial ecosystems, but was comparable to phenological studies of other marine organisms, indicating that climate change may lead to a rapid reorganization of seasonal patterns in marine ichthyoplankton communities. Compared with historical changes, IPCC models appear to underestimate future projections of phenological change. In contrast to other ecosystems, both sustained advances and delays in phenology were observed in the CCE, reflecting the opposing influences of temperature and upwelling on phenology. Similar changes are likely to occur in other EBCU ecosystems if global

warming alters the seasonal intensity of upwelling. Because zooplankton phenology did not change in synchrony with the seasonal larval abundance of many fishes, these species could be subject to increased mismatches with their prey.

Materials and Methods

Ichthyoplankton samples were collected by CalCOFI along six transects (lines 76.7–93.3), which extended offshore of California between San Diego (33.0°N) and north of Point Conception (35.1°N). During each cruise, 66 stations were typically surveyed with oblique tows of bongo or ring nets (Fig. S1). Kramer et al. described methods used to catch, preserve, and identify ichthyoplankton (41). Throughout this time series, two modifications were made to sampling methods. First, tow depth was increased from 140 to 210 m in 1969. This event was accompanied by a change from using a net with 550- μ m silk mesh to a 505- μ m mesh. Second, the 1.0-m ring net initially used to sample ichthyoplankton was replaced with a 0.71-m bongo net in 1977 (80). Although slightly more larvae >6.75 mm total length were caught with the bongo net, large larvae made up such a small fraction of total abundance that there was not a significant difference between the number of larvae captured by the two nets (54). The change in tow depth also had a minimal effect on abundance because most larvae resided at depths <125 m (56, 81). Data were averaged decadal to develop monthly time series.

Only taxa identified to species level were included in this analysis to avoid confounding changes in species composition with shifts in phenology (7). Cumulative rank occurrence of ichthyoplankton taxa sampled between 1951 and 1998 was used to select species for inclusion in this study. The 90th percentile of the cumulative rank occurrence curve included 43 species, which were examined here. Due to improvements over time in identifying larval fishes, some taxa were not consistently identified during the first two decades of CalCOFI. As a result, insufficient data were available to analyze three species during the 1950s (*Argyropelecus sladeni*, *Lestidiops ringens*, *Oxyjulis californica*) and an additional three species during both the 1950s and 1960s (*Cyclothone signata*, *Danaphos oculatus*, *Melamphaes lugubris*).

Some species exhibited peaks in larval abundance during both spring and fall. In such cases, three criteria were used to determine whether a species with several seasonal peaks in abundance should be classified as having multiple phenophases. First, peaks in abundance must be separated by ≥ 3 mo. Second, the amplitude of each peak must be >50% of the annual range of larval abundance. Third, the dip in abundance between peaks must encompass >35% of the annual range. Mean seasonal abundance from the full time series was used to evaluate these criteria.

CT was used as the main measure of phenology in this study because it accounted for data collected during all seasons, can be calculated from a time series with a monthly resolution, and is a conservative metric that is unlikely to overestimate the rate of phenological change (61). The CT of each phenophase was calculated with the following equation (7):

$$CT = \frac{\sum_{i=1}^{12} (m_i \times a_i)}{\sum_{i=1}^{12} a_i}$$

where a_i was mean abundance of larvae in month m_i , such that $i = 1, 2, \dots, 12$ denoted January, February, ... December. For species with two phenophases, each year was divided into two 6-mo periods (typically January–June and July–December) for which the CT was computed. For species whose abundance increased between December and January, the CT was calculated so that $i = 1$ referred to the month when the seasonal rise in abundance started. Phenological anomalies were calculated relative to the mean of the CTs from the six decades examined to ensure that each decade was weighted equally regardless of whether fish abundance was high or low. An anomaly of 1 indicated a 1-mo change in phenology. To convert monthly anomalies into daily values, they were multiplied by 30.44.

Two approaches were applied to examine assemblage-wide trends in phenology. First, a linear regression was performed in which decadal CT anomalies of all phenophases were regressed against year. Here and elsewhere phenophases were treated as replicates for examining assemblage-wide patterns. The slope of this regression was used to estimate the mean rate of change in days per decade. First-order, autoregressive (AR1) terms were incorporated into this regression and all other temporal analyses when AR1 terms were significant at $P < 0.05$ and the inclusion of AR1 minimized the Akaike information criterion (AIC). Second, PCA was used to reduce the dimensionality of the community composition data while further assessing changes within the assemblage's first mode of variability. The CT of each phenophase was standardized by its mean and SD to give equal weight to all phenophases in the PCA (82). Species with missing data from the 1950s and

1960s were excluded from this analysis, so that the correlation matrix for the PCA would be complete.

Phenophases were classified into three groups based on correlations between CT anomalies and year. Phenophases with $r \leq -0.5$ were categorized as exhibiting earlier phenology; correlations between -0.5 and 0.5 indicated no long-term, linear change; and phenophases with $r \geq 0.5$ were classified as displaying later phenology. Note that $|r| \geq 0.5$ does not correspond to a specific threshold of statistical significance, but is instead intended to provide a liberal indication of the direction of phenological change. As was done for the assemblage-wide mean, the slope of regressions between CT anomalies and year (plus an AR1 term, if statistically significant) were used to estimate rates of change for phenophases in each group. Monte Carlo simulation was used to determine the CIs for long-term changes in CT (SI Text and Tables S3 and S4). SI Text and Fig. S2 also include an evaluation of potential bias due to missing data from May, September, and December during the 2000s.

This study examined how three basin-scale climate oscillations, three local oceanic variables, and nine ecological characteristics affected the phenology of larval fishes. Details of each analysis are described in the SI Text. Climate indices (e.g., ENSO, PDO, NPGO) were partitioned into a set of three categorical variables for use in ANOVA. Data on each fish phenophase were aggregated into bins corresponding to the categorical phases of the climate indices. CT anomalies of phenophases were then calculated relative to each climate index phase. Lilliefors' test, Bartlett's test, and plots of residuals were used to ascertain whether the ANOVA assumptions of normality, homoscedasticity, and independence of residuals were met (82). If an ANOVA indicated a significant climate effect, Tukey–Kramer multiple comparison tests were used post hoc to determine which climate phases exhibited significant phenological differences. The CT of local oceanic variables (e.g., CalCOFI SST, CalCOFI zooplankton volume, Bakun upwelling index) was calculated on a decadal basis to identify long-term trends and fluctuations in the seasonality of these variables. The Bakun upwelling index overestimates offshore transport across the CalCOFI region due to a discontinuity in the atmospheric pressure gradient related to the presence of coastal mountains (83). However, because this index has been measured consistently over time, this bias seems unlikely to result in a spurious trend in phenology (SI Text). Forward stepwise, multiple regressions were performed to investigate the relationship between the CT anomalies of oceanic variables and the CT anomalies of fishes in each phenology group. Because several ecological characteristics of fishes were correlated, PCA was used to identify major modes of variability among the ecological characteristics. The first two PCs of the ecological traits were then compared with changes in fish CT with Pearson correlation tests. SI Text and Table S6 also describe additional tests investigating interactions between trends in phenology, shifts in the geographic range of species, and decadal changes in larval abundance.

Projections of 21st century changes in phenology were based on a multi-model dataset developed during phase 3 of the World Climate Research Program's (WCRP's) Coupled Model Intercomparison Project (CMIP3) (www.pcmdi.llnl.gov/ipcc/about_ipcc.php). Of the 25 models included in CMIP3, 12 were selected for examination due to their ability to replicate the spatial signature of the PDO with a correlation of $r \geq 0.7$ (64). Data on SST and coastal upwelling derived from meridional wind speed were extracted from CMIP3 to serve as predictors of changes in fish phenology. Thirty simulations from the ensemble of 12 models contained SST data. Wind speed was available from 26 simulations. Monthly data from the SRES A1B scenario were examined during 2000–2099 and averaged over 30–35°N and 117–125°W. Projections of seasonal upwelling were unaffected by whether winds were extracted from the full study area or solely the area west of Point Conception, where upwelling is maximal off southern California (17). To calculate the volume of offshore Ekman transport (Q_x), wind stress (τ) was first computed with the formula: $\tau_y = \rho_{air} C_d v^2$, where v is meridional wind speed at an atmospheric pressure of 100 kPa, ρ_{air} is the density of air, and C_d is the drag coefficient (1.2×10^{-3}) (84). Q_x was calculated from the equation $Q_x = \tau_y \times (\rho f)^{-1}$, where ρ is the density of seawater and f is the Coriolis parameter (85). A mean surface water density of 1,024.5 kg/m³ and an f of $7.83 \times 10^{-5} \text{ s}^{-1}$ (based on a mean latitude of 32.5°N) were assumed for the CalCOFI region. Q_x was multiplied by 100 because the Bakun upwelling index examined Ekman transport over 100 m of coastline (83). CT anomalies of SST and upwelling were calculated decadal from CMIP3 data using the 2000s as a baseline for computing anomalies. The regressions between CT anomalies of local, oceanic variables, and species with earlier or delayed phenology (Fig. 3 and Table S5) were used to make projections of future phenological change. Because zooplankton volume is not included as a state variable in CMIP3 models, a CT anomaly of zero was assumed for this term in the regression for earlier species. Multimodel means of phenological change were calculated such that each model received equal weight regardless of the number of simulations run per model.

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